



Development Support Document
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Hexavalent Chromium Oral Reference Dose

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PROPOSED

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TEXAS COMMISSION ON ENVIRONMENTAL QUALITY

1 **Executive Summary**

2 ***Background***

3 In recent years, a great deal of new research has been conducted specifically to generate data to
4 better inform the mode of action (MOA) analysis for hexavalent chromium-induced
5 carcinogenesis due to chronic oral exposure and to improve the extrapolation of rodent oral study
6 results to humans (e.g., Thompson et al. 2011a, 2011b, 2012a, 2012b, 2012c, 2013a; Kirman et
7 al. 2012, 2013; Proctor et al. 2012; Kopec et al. 2012a, 2012b; O'Brien et al. 2013; Suh et al.
8 2014; Thompson et al. 2015a, 2015b, 2015c). These research project data have been thoroughly
9 evaluated to gain critical scientific understanding and insight into key areas of carcinogenic
10 dose-response assessment for hexavalent chromium (CrVI) such as:

- 11 • The carcinogenic MOA (i.e., key events) operating in relevant rodent studies (e.g., NTP
12 2008);
- 13 • CrVI toxicokinetics following oral exposure (e.g., dose-dependent differences in target
14 tissue absorption); and
- 15 • Data-informed, biologically-plausible expectations about potential low-dose risk.

16 The current, significantly greater scientific understanding of these issues (especially considering
17 the previous lack of sufficient relevant data and understanding just 5-6 years ago) is of
18 paramount importance considering the substantial regulatory challenge of extrapolating high oral
19 dose results from laboratory animal studies to environmentally-relevant human doses that are
20 orders of magnitude lower in a meaningful, toxicologically-predictive manner (e.g., the mouse
21 dose at the lowest water concentration used in NTP 2008 is about 74,000 times higher than the
22 approximate human dose corresponding to the 35-city geometric mean drinking water
23 concentration reported in EWG 2010).

24 ***TCEQ Scientific, Peer-Reviewed Publications***

25 In addition to the numerous studies published as part of the CrVI MOA research project, TCEQ
26 staff have independently and critically evaluated the relevant published study data (and
27 supplemental data) and its implications for the carcinogenic dose-response assessment of oral
28 exposure to CrVI. As part of that scientific endeavor, three manuscripts have been published in a
29 scientific peer-reviewed journal (Haney 2015a, 2015b, 2015c). The last of these scientific, open-
30 access articles (Haney 2015c) considers both the non-linear, non-threshold approach as well as
31 the threshold (i.e., reference dose) approach prior to conducting a weight-of-evidence (WOE)
32 analysis of available MOA data. The WOE indicates that cytotoxicity-induced regenerative
33 hyperplasia is indubitably the most scientifically well-supported MOA. Health Canada (2015)
34 concurs that confidence in a cytotoxic MOA is high (and evidence for a mutagenic MOA is
35 weak). More specifically, compensatory crypt enterocyte hyperplasia induced by chronic villous
36 toxicity should be considered as a required (but not always sufficient) key event in CrVI-induced
37 intestinal tumorigenesis. Consequently, the reference dose (RfD) approach is the most

1 scientifically-defensible approach based on the WOE of available MOA information and
2 analyses conducted for the most scientifically-supported MOA and should be adopted for
3 assessing the potential intestinal carcinogenicity of oral exposure to CrVI (Haney 2015c).

4 ***RfD Derivation***

5 The RfD derived in Haney (2015c) will be adopted by the TCEQ. The RfD of 0.0031 mg
6 CrVI/kg-day was derived to protect against cytotoxicity-induced regenerative hyperplasia as a
7 key precursor carcinogenic MOA event. Briefly, the duodenum was selected as the critical
8 mouse target tissue for benchmark dose (BMD) analysis since diffuse hyperplasia has a strong,
9 well-defined dose-response relationship in the mouse duodenum. This is consistent with both
10 significant tissue absorption of CrVI by the duodenum and the duodenum as the most
11 tumorigenically responsive tissue. The incidence of diffuse hyperplasia in the duodenum of
12 female mice was used for BMD modeling since: (1) Statistical analyses did not reveal
13 differences between male and female mice in hyperplastic or tumorigenic response to CrVI
14 exposure (Thompson et al. 2013b); (2) The dose-response for diffuse hyperplasia in female mice
15 is strong and more monotonic than that in male mice (see Tables C4 and D4 of NTP 2008); and
16 (3) Importantly, the water concentrations used in NTP (2008) for female mice correspond to
17 those used in Kirman et al. (2012) to determine added chromium (Cr) concentrations in mouse
18 target tissues due to CrVI oral exposure, which is a useful internal dose metric for BMD
19 modeling. Accordingly, the incidence of diffuse hyperplasia in the duodenum of female mice
20 from NTP (2008) along with the duodenum tissue concentrations (added mg Cr/kg tissue)
21 reported in Kirman et al. (2012) were used for BMD modeling. A benchmark response (BMR) of
22 10% was used so that the BMD and 95% lower confidence limit on the BMD (BMDL) would be
23 calculated at a BMR that did not extrapolate farther than necessary below the range of the data.

24 The Log-Logistic and Dichotomous-Hill models provided adequate and almost identical fits to
25 the mouse data with a goodness-of-fit p value >0.1 , lowest Akaike Information Criterion, and
26 scaled residuals $<|2|$. The mouse BMD₁₀ value was 1.83 added mg Cr/kg tissue for both models
27 using mean added mg Cr/kg tissue as the internal dose metric. The average mouse BMDL₁₀ of
28 1.39 added mg Cr/kg tissue (based on individual model values of 1.37 and 1.41 added mg Cr/kg
29 tissue) was used as the point of departure (POD) for diffuse hyperplasia in the duodenum for
30 derivation of the RfD.

31 The mouse POD of 1.39 added mg Cr/kg duodenum tissue was converted to a corresponding oral
32 dose based on the relationship between duodenum tissue concentration (mean mg Cr/kg tissue)
33 and oral dose (mg CrVI/kg-day) that was modeled previously (Haney 2015a). The POD falls
34 between two of the tissue concentrations modeled, and is similar to one of the modeled
35 concentrations (1.5 mg Cr/kg tissue) where the estimated and observed values showed excellent
36 agreement (i.e., the scaled residual was 0.421, well below $|2|$), which increases confidence in the
37 oral exposure estimate corresponding to the target tissue dose POD. A mouse oral dose of 0.31
38 mg CrVI/kg-day was estimated to correspond to the POD duodenum tissue concentration. Note
39 that the application of an animal-to-human uncertainty factor (UF_A) to this mouse POD

1 ultimately results in a value (0.031 mg/kg-day) that is below the lower end of the range of
2 average human equivalent doses (HED values of 0.05-0.1mg/kg-day) cited in a recent USEPA
3 CrVI PBPK study (Sasso and Schlosser 2015), practically identical to the more conservative
4 HED of 0.028 mg/kg-day (pH = 5) based on a similar evaluation (e.g., using the BMDL₁₀ for
5 diffuse epithelial hyperplasia), and is 4.5-fold lower than the HED of 0.14 mg/kg-day (pH = 2.5)
6 based on the similar evaluation (see Table 1 of Sasso and Schlosser 2015).

7 Dividing this mouse oral POD (0.31 mg CrVI/kg-day) by the same uncertainty factors (UF_A=10,
8 UF_H=10, UF_D=1) as used in USEPA (2010) results in an RfD of 0.0031 mg CrVI/kg-day.

9 RfD = 3.1E-03 mg CrVI/kg-day

10 ***Comparison of TCEQ RfD with Other Published RfDs***

11 The TCEQ-derived RfD is somewhat more conservative than, but shows remarkable agreement
12 with, a previously published RfD (0.006 mg CrVI/kg-day; Thompson et al. 2013b) as well as
13 Health Canada's Tolerable Daily Intake (TDI of 0.0044 mg CrVI/kg-day; Health Canada 2015).
14 It also happens to correspond to the approximate human dose at the federal MCL for Cr (e.g., 0.1
15 mg/L × 2.5 L/day × 1/80 kg = 0.0031 mg/kg-day), and is considered protective of both the
16 potential carcinogenic and noncarcinogenic effects of oral CrVI exposure. Additional details
17 pertaining to the RfD derivation may be found in the open access article ([Haney 2015c](#)).

18 **References**

19 Environmental Working Group (EWG). 2010. Chromium-6 in US tap water. Environmental
20 Working Group, December 20, 2010.

21 Haney J. 2015a. Use of dose-dependent absorption into target tissues to more accurately predict
22 cancer risk at low oral doses of hexavalent chromium. *Regul Toxicol Pharmacol* 71, 93-
23 100.

24 Haney J. 2015b. Implications of dose-dependent target tissue absorption for linear and non-
25 linear/threshold approaches in development of a cancer-based oral toxicity factor for
26 hexavalent chromium. *Regul Toxicol Pharmacol* 72, 194-201.

27 Haney J. 2015c. Consideration of non-linear, non-threshold and threshold approaches for
28 assessing the carcinogenicity of oral exposure to hexavalent chromium. *Regul Toxicol*
29 *Pharmacol* 73, 834-852. Available at:
30 <http://www.sciencedirect.com/science/article/pii/S0273230015300957>.

31 Health Canada. 2015. Chromium in drinking water (document for public consultation). Prepared
32 by the Federal-Provincial-Territorial Committee on Drinking Water, Health Canada.

- 1 Kirman C, Hays S, Aylward L, et al. 2012. Physiologically based pharmacokinetic model for rats
2 and mice orally exposed to chromium. *Chem Biol Interact* 200, 45-64.
- 3 Kirman C, Aylward L, Suh M, et al. 2013. Physiologically based pharmacokinetic model for
4 humans orally exposed to chromium. *Chem Biol Interact* 204, 13-27.
- 5 Kopec A, Kim S, Forgacs A, et al. 2012a. Genome-wide gene expression effects in B6C3F1
6 mouse intestinal epithelia following 7 and 90 days of exposure to hexavalent chromium
7 in drinking water. *Toxicol Appl Pharmacol* 259, 13-26.
- 8 Kopec A, Thompson C, Kim S, et al. 2012b. Comparative toxicogenomic analysis of oral Cr(VI)
9 exposure effects in rat and mouse small intestinal epithelium. *Toxicol Appl Pharmacol*
10 262, 124-138.
- 11 National Toxicology Program (NTP). 2008. National Toxicology Program technical report on
12 the toxicology and carcinogenesis studies of sodium dichromate dihydrate (CAS No.
13 7789-12-0) in F344/N rats and B6C3F1 mice (drinking water studies). NTP Toxicity
14 Report 546. NIH Publication No. 08-5887.
- 15 O'Brien T, Ding H, Suh M, et al. 2013. Assessment of K-Ras mutant frequency and
16 micronucleus incidence in the mouse duodenum following 90-days of exposure to Cr(VI)
17 in drinking water. *Mutat Res* 754, 15-21.
- 18 Proctor D, Suh M, Aylward L, et al. 2012. Hexavalent chromium reduction kinetics in rodent
19 stomach contents. *Chemosphere* 89, 487-493.
- 20 Sasso A, and Schlosser P. 2015. An evaluation of in vivo models for toxicokinetics of hexavalent
21 chromium in the stomach. *Toxicol Appl Pharmacol* 287, 293-298.
- 22 Suh M, Thompson C, Kirman C, et al. 2014. High concentrations of hexavalent chromium in
23 drinking water alter iron homeostasis in F344 rats and B6C3F1 mice. *Food Chem Toxicol*
24 65, 381-388.
- 25 Thompson C, Proctor D, Haws L, et al. 2011a. Investigation of the mode of action underlying the
26 tumorigenic response induced in B6C3F1 mice exposed orally to hexavalent chromium.
27 *Toxicol Sci* 123, 58-70.
- 28 Thompson C, Haws L, Harris M, et al. 2011b. Application of the US EPA mode of action
29 framework for purposes of guiding future research: a case study involving the oral
30 carcinogenicity of hexavalent chromium. *Toxicol Sci* 119, 20-40.

- 1 Thompson C, Proctor D, Suh M, et al. 2012a. Comparison of the effects of hexavalent chromium
2 in the alimentary canal of F344 rats and B6C3F1 mice following exposure in drinking
3 water: implications for carcinogenic modes of action. *Toxicol Sci* 125, 79-90.
- 4 Thompson C, Hixon J, Proctor D, et al. 2012b. Assessment of genotoxic potential of Cr(VI) in
5 the mouse duodenum: an in silico comparison with mutagenic and nonmutagenic
6 carcinogens across tissues. *Regul Toxicol Pharmacol* 64, 68-76.
- 7 Thompson C, Fedorov Y, Brown D, et al. 2012c. Assessment of Cr(VI)-induced cytotoxicity and
8 genotoxicity using high content analysis. *PLoS One*, 7, e42720.
- 9 Thompson C, Proctor D, Suh M, et al. 2013a. Assessment of the mode of action underlying
10 development of rodent small intestinal tumors following oral exposure to hexavalent and
11 relevance to humans. *Crit Rev Toxicol* 43, 244-274.
- 12 Thompson C, Kirman C, Proctor D, et al. 2013b. A chronic oral reference dose for hexavalent
13 chromium-induced intestinal cancer. *J Appl Toxicol* 34, 525-536.
- 14 Thompson C, Seiter J, Chappell M, et al. 2015a. Synchrotron-based imaging of chromium and
15 gamma-H2AX immunostaining in the duodenum following repeated exposure to Cr(VI)
16 in drinking water. *Toxicol Sci* 143, 16-25.
- 17 Thompson C, Young R, Suh M, et al. 2015b. Assessment of the mutagenic potential of Cr(VI) in
18 the oral mucosa of big blue® transgenic f344 rats. *Environ Mol Mutagen* 56, 621-628.
- 19 Thompson C, Wolf J, Elbekai R, et al. 2015c. Duodenal crypt health following exposure to
20 Cr(VI): Micronucleus scoring, γ -H2AX immunostaining, and synchrotron X-ray
21 fluorescence microscopy. *Mutat Res: Genet Toxicol Environ Mutagen* 789-790, 61-66.
- 22 United States Environmental Protection Agency (USEPA). 2010. Toxicological Review of
23 Hexavalent Chromium In Support of Summary Information on the Integrated Risk
24 Information System (IRIS). EPA/635/R-10/004A. US Environmental Protection Agency,
25 Washington, DC.